

Global fine-scale changes in ambient NO₂ during COVID-19 lockdowns

Cooper, Matthew J.; Martin, Randall V.; Hammer, Melanie S.; Levelt, Pieternel F.; Veefkind, Pepijn; Lamsal, Lok N.; Krotkov, Nickolay A.; Brook, Jeffrey R.; McLinden, Chris A.

DOI

[10.1038/s41586-021-04229-0](https://doi.org/10.1038/s41586-021-04229-0)

Publication date

2022

Document Version

Final published version

Published in

Nature

Citation (APA)

Cooper, M. J., Martin, R. V., Hammer, M. S., Levelt, P. F., Veefkind, P., Lamsal, L. N., Krotkov, N. A., Brook, J. R., & McLinden, C. A. (2022). Global fine-scale changes in ambient NO₂ during COVID-19 lockdowns. *Nature*, 601(7893), 380-387. <https://doi.org/10.1038/s41586-021-04229-0>²

Important note

To cite this publication, please use the final published version (if applicable).
Please check the document version above.

Copyright

Other than for strictly personal use, it is not permitted to download, forward or distribute the text or part of it, without the consent of the author(s) and/or copyright holder(s), unless the work is under an open content license such as Creative Commons.

Takedown policy

Please contact us and provide details if you believe this document breaches copyrights.
We will remove access to the work immediately and investigate your claim.

Global fine-scale changes in ambient NO₂ during COVID-19 lockdowns

<https://doi.org/10.1038/s41586-021-04229-0>

Received: 9 February 2021

Accepted: 11 November 2021

Published online: 19 January 2022

Open access

 Check for updates

Matthew J. Cooper^{1,2✉}, Randall V. Martin^{1,2,3}, Melanie S. Hammer^{1,2}, Pieternel F. Levelt^{4,5,6}, Pepijn Veefkind^{4,7}, Lok N. Lamsal^{8,9}, Nickolay A. Krotkov⁸, Jeffrey R. Brook^{10,11} & Chris A. McLinden¹²

Nitrogen dioxide (NO₂) is an important contributor to air pollution and can adversely affect human health^{1–9}. A decrease in NO₂ concentrations has been reported as a result of lockdown measures to reduce the spread of COVID-19^{10–20}. Questions remain, however, regarding the relationship of satellite-derived atmospheric column NO₂ data with health-relevant ambient ground-level concentrations, and the representativeness of limited ground-based monitoring data for global assessment. Here we derive spatially resolved, global ground-level NO₂ concentrations from NO₂ column densities observed by the TROPOMI satellite instrument at sufficiently fine resolution (approximately one kilometre) to allow assessment of individual cities during COVID-19 lockdowns in 2020 compared to 2019. We apply these estimates to quantify NO₂ changes in more than 200 cities, including 65 cities without available ground monitoring, largely in lower-income regions. Mean country-level population-weighted NO₂ concentrations are 29% ± 3% lower in countries with strict lockdown conditions than in those without. Relative to long-term trends, NO₂ decreases during COVID-19 lockdowns exceed recent Ozone Monitoring Instrument (OMI)-derived year-to-year decreases from emission controls, comparable to 15 ± 4 years of reductions globally. Our case studies indicate that the sensitivity of NO₂ to lockdowns varies by country and emissions sector, demonstrating the critical need for spatially resolved observational information provided by these satellite-derived surface concentration estimates.

Nitrogen dioxide (NO₂) is an important contributor to air pollution as a primary pollutant and as a precursor to ozone and fine particulate matter production. Human exposure to elevated NO₂ concentrations is associated with a range of adverse outcomes such as respiratory infections^{2–4}, increases in asthma incidence^{5,6}, lung cancer⁷ and overall mortality^{8,9}. NO₂ observations indicate air quality relationships with combustion sources of pollution such as transportation^{6,21}. Initial investigations found substantial decreases in the atmospheric NO₂ column from satellite observations^{10–16} and in ambient NO₂ concentrations from ground-based monitoring^{17–20} during lockdowns enacted to reduce the spread of COVID-19. However, questions remain about the relationship of atmospheric columns with health- and policy-relevant ambient ground-level concentrations, and about the representativeness of sparse ground-based monitoring for broad assessment. Thus, there is need to relate satellite observations of NO₂ columns to ground-level concentrations. It is also important to consider the effect of meteorology on recent NO₂ changes²² and to quantify NO₂ changes due to COVID-19 interventions in the context of longer-term trends²³. Furthermore, air quality monitoring sites tend to be preferentially located in higher-income regions, raising questions

about how NO₂ changed in lower-income regions where larger numbers of potentially susceptible people reside. Estimates of changes in ground-level NO₂ concentrations derived from satellite remote sensing would fill gaps between ground-based monitors, offer valuable information in regions with sparse monitoring, and more clearly connect satellite observations with ground-level ambient air quality.

Previous satellite-derived estimates of ground-level NO₂ used information on the vertical distribution of NO₂ from a chemical transport model to relate satellite NO₂ column densities to ground-level concentrations^{24–26}. Recent work improved upon this technique by allowing the satellite column densities to constrain the vertical profile shape, allowing for more accurate representation of sub-model-grid variability, reducing the sensitivity to model resolution and simulated profile shape errors, and improving agreement between the satellite-derived ground-level concentrations and in situ monitoring data²⁷. Applying this technique to examine changes in NO₂ during lockdowns bridges the gap between previous studies focusing on either ground monitors or satellite column densities, thus providing a more complete and reliable picture of the changes in exposure.

¹Department of Physics and Atmospheric Science, Dalhousie University, Halifax, Nova Scotia, Canada. ²Department of Energy, Environmental & Chemical Engineering, Washington University in St. Louis, St. Louis, MO, USA. ³Harvard-Smithsonian Center for Astrophysics, Cambridge, MA, USA. ⁴Royal Netherlands Meteorological Institute (KNMI), De Bilt, Netherlands. ⁵University of Technology Delft, Delft, Netherlands. ⁶National Center for Atmospheric Research, Boulder, CO, USA. ⁷Department of Geoscience and Remote Sensing, Delft University of Technology, Delft, Netherlands. ⁸NASA Goddard Space Flight Center, Greenbelt, MD, USA. ⁹Universities Space Research Association, Columbia, MD, USA. ¹⁰Dalla Lana School of Public Health, University of Toronto, Toronto, Ontario, Canada. ¹¹Department of Chemical Engineering and Applied Chemistry, University of Toronto, Toronto, Ontario, Canada. ¹²Environment and Climate Change Canada, Toronto, Ontario, Canada. ✉e-mail: cooperm2@dal.ca

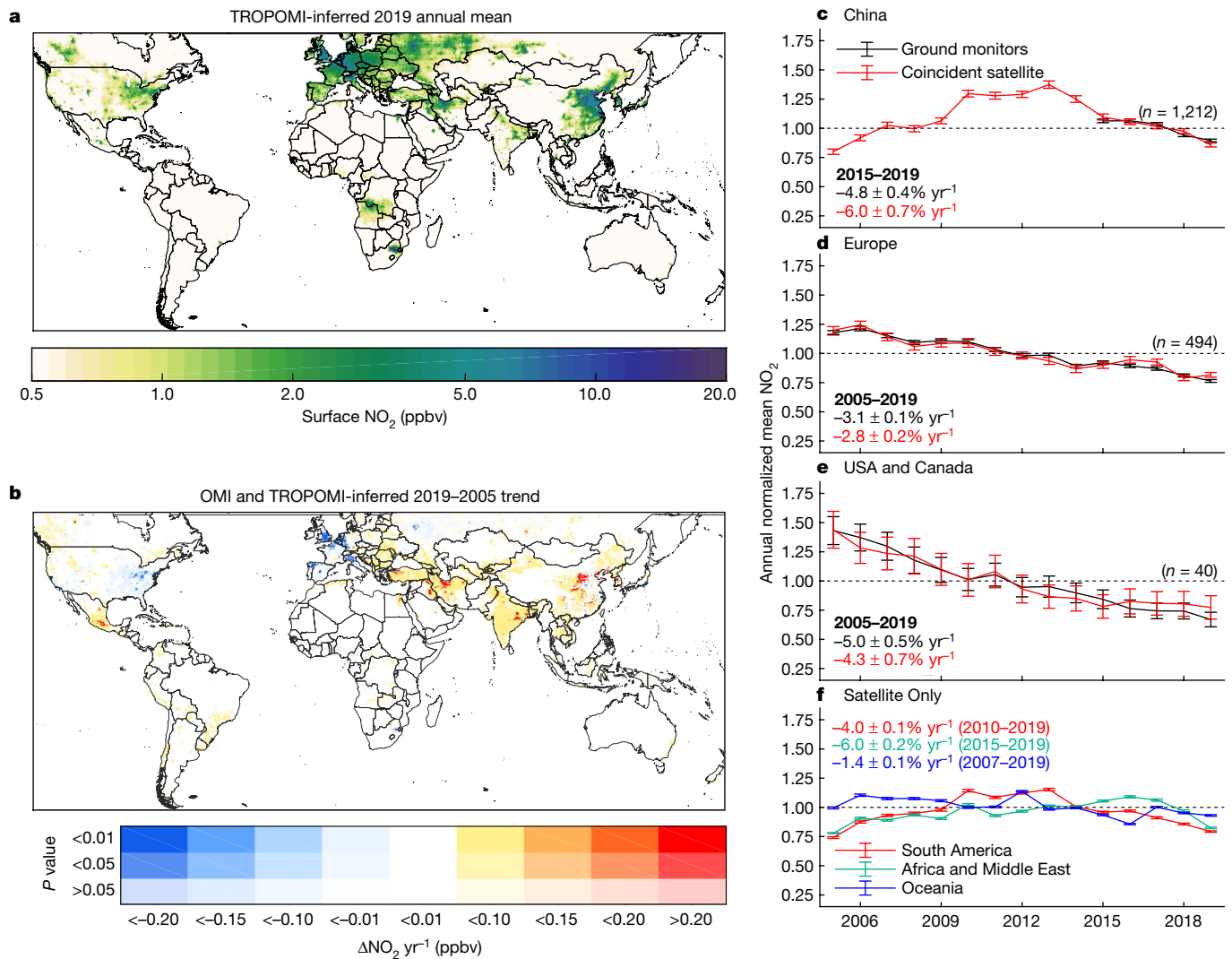


Fig. 1 | Satellite-derived ground-level NO₂ concentrations. **a**, TROPOMI-derived 2019 annual mean ground-level NO₂ concentrations at approximately 1 × 1 km² resolution. **b**, Trend in OMI and TROPOMI-derived annual mean ground-level concentrations from 2005–2019. The colour intensity represents the statistical significance of the trend. **c–e**, Population-weighted mean NO₂ from ground monitors and from satellite-derived NO₂ sampled at ground-monitor locations in China (**c**), Europe (**d**) and North America (**e**), normalized by the mean concentration during the period where ground-monitor data are available. The black (ground-derived) and red (satellite-derived) values give the

trends for the period where ground-monitor data are available. Only monitors with data available over the entire time period are included. Error bars represent population-weighted standard deviations. **f**, Population-weighted mean satellite-inferred ground-level NO₂ concentrations in South America, Africa and the Middle East, and Oceania. Trends during the given time periods are given at top. Time periods were chosen to reflect the most recent years where a consistent trend is observed. Error bars represent uncertainties in population-weighted means using a bootstrapping method.

Since 2005, the gold standard for satellite NO₂ observations has been the Ozone Monitoring Instrument (OMI) on board NASA's Earth Observing System Aura satellite^{28,29}. The newest remote sensing spectrometer, the European Space Agency's TROPospheric Monitoring Instrument (TROPOMI)³⁰ on the Copernicus Sentinel 5p satellite, has been providing NO₂ observations with finer spatial resolution and higher instrument sensitivity since 2018. These attributes allow the generation of TROPOMI NO₂ maps at 100 times finer resolution (approximately 1 × 1 km²) with a one-month averaging period^{31,32}, an improvement over the spatial and temporal averaging needed for accurate OMI maps (typically approximately 10 × 10 km² over one year)²⁴. Concurrently, the excellent stability of the OMI instrument over the last 15 years provides an ideal dataset for long-term trend analysis^{28,33} that offers context for recent TROPOMI data.

Lockdown restrictions act as an experiment about the efficacy of activity reductions on mitigating air pollution. The Oxford COVID-19

Government Response Tracker (OxCGRT, <https://www.bsg.ox.ac.uk/research/research-projects/coronavirus-government-response-tracker/#data>) has been monitoring government-imposed restrictions, and studies have indicated that NO₂ decreases were larger for cities in countries with strict lockdowns³⁴. However, there is limited information on lockdown stringency on sub-national levels or on how various emission sectors respond to lockdowns. An observation-based metric for lockdown intensity could provide useful information for examining lockdowns on city-level scales or for examining the effects on air quality that are associated with lockdowns in different emission sectors.

Here we leverage the high spatial resolution of TROPOMI to infer global ground-level NO₂ estimates at, to our knowledge, an unprecedented spatial resolution sufficient to assess individual cities worldwide, and to examine changes in ground-level NO₂ occurring during COVID-19 lockdowns from January–June 2020. Case studies presented here demonstrate how the satellite-based estimates provide

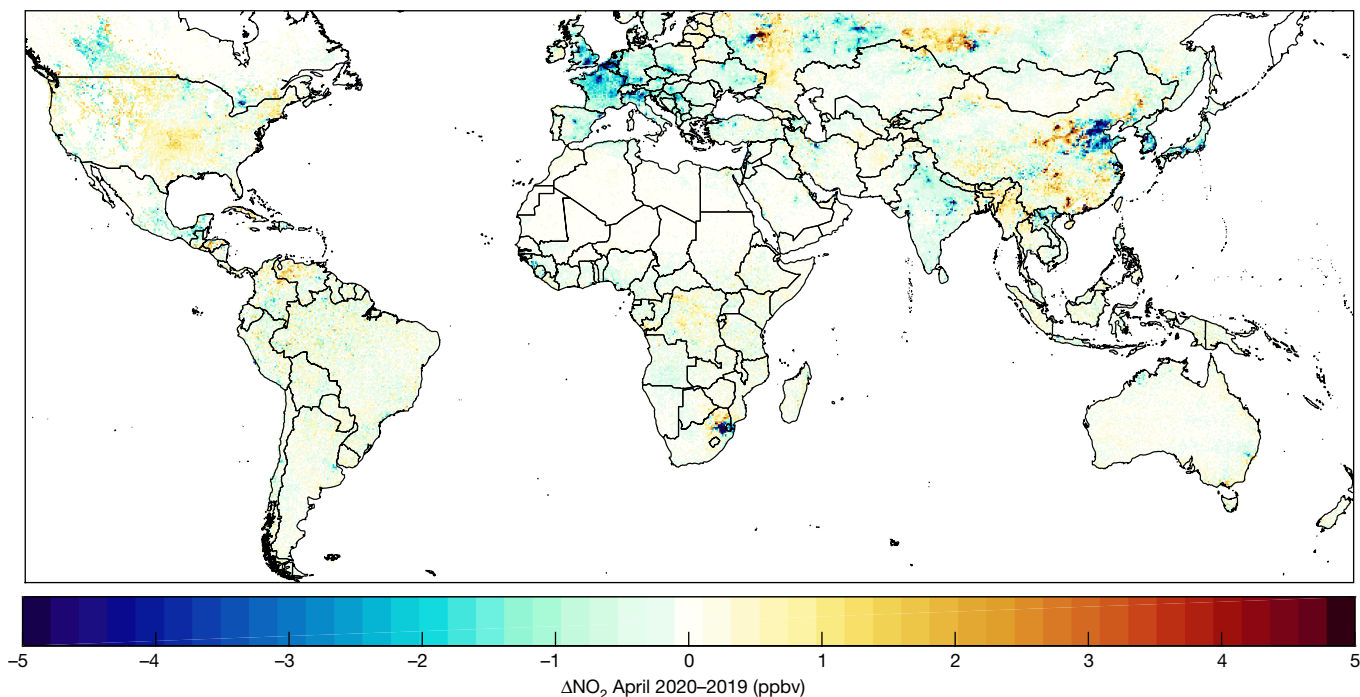


Fig. 2 | Differences in April mean ground-level NO₂ from 2020 to 2019. Concentrations derived using TROPOMI observations gridded at approximately 1 × 1-km² resolution.

information on important spatial variability in lockdown-driven NO₂ changes, and in the NO₂ response to lockdowns in various emissions sectors. We also use TROPOMI to provide fine-scale structure to the long-term record of OMI observations (2005–2019), which provides an opportunity to examine trends in ground-level NO₂ over the last 15 years to provide context for the recent changes.

Global NO₂ concentrations and trends

Global annual mean TROPOMI-derived ground-level NO₂ concentrations for 2019 provide an initial baseline (Fig. 1). The excellent resolution (1 × 1 km²) of ground-level NO₂ concentrations reveal pronounced heterogeneity (Supplementary Figs. 1–7). NO₂ enhancements are apparent over urban and industrial regions. TROPOMI-derived ground-level concentrations exhibit consistency with in situ observations ($r = 0.71$, $N = 3,977$, in situ versus satellite slope = 0.97 ± 0.02), as shown in Supplementary Fig. 8. Neglecting the spatial and temporal variability in the NO₂ column-to-surface relationship degrades the consistency with ground monitors (slope = 0.78 ± 0.01), demonstrating the importance of relating satellite columns to surface concentrations for exposure assessment.

Examination of long-term changes in air pollution offers context for changes during COVID-19 lockdowns (Fig. 1, Supplementary Figs. 1–7). Satellite-derived NO₂ concentrations decreased from 2005–2019 in urban areas across most of the USA and Europe, eastern China, Japan, and near Johannesburg, South Africa, largely reflecting emission controls on vehicles and power generation. NO₂ increases are observed in Mexico, the Alberta oil sands region in northern Canada, and throughout the Balkan peninsula, central and northern China, India and the Middle East, broadly consistent with reported trends in ground-monitor data^{35–37}. Trends in China can be separated into three regimes: ground-level concentrations increased in China from 2005–2010, plateaued from 2010–2013, and decreased from 2013–2019. This change was driven by stricter vehicle and power generation emission standards³⁸ and is consistent with observed changes in NO₂ columns^{39,40}. Similarly, concentrations increased in urban and industrial areas of

South America from 2005–2010, and in South Africa and the Middle East from 2005–2015, and decreased in more recent years. Maps of trends in these regions for these time periods are shown in Supplementary Fig. 9. Concentrations in India increased across both time periods owing to increasing coal-powered electricity demands and growing industrial emissions⁴¹. Trends in population-weighted NO₂ concentrations, used to represent population NO₂ exposure, were calculated using ground monitors and coincidentally sampled satellite observations in North America, Europe and China. Satellite-derived concentrations exhibit decreasing trends ($-2.8 \pm 0.2\% \text{ yr}^{-1}$ in Europe 2005–2019, $-4.3 \pm 0.7\% \text{ yr}^{-1}$ in North America 2005–2019, and $-6.0 \pm 0.7\% \text{ yr}^{-1}$ in China 2015–2019) that agree well with trends in the ground-monitor data (within $0.7\% \text{ yr}^{-1}$ in North America, $0.3\% \text{ yr}^{-1}$ in Europe, and $1.2\% \text{ yr}^{-1}$ in China).

Regional NO₂ changes during lockdowns

Figure 2 shows the April 2020 to April 2019 difference between mean ground-level NO₂ concentrations derived from TROPOMI observations. NO₂ concentrations are lower in most regions in 2020 than in 2019, particularly over urban areas, with global population-weighted mean concentrations decreasing by 16% in 2020 relative to 2019. Fig. 3 shows regional maps focusing on the month with the largest change in population-weighted regional mean concentration for each region, with an additional period included for China, as lockdown restrictions occurred earlier than in other countries. Regional population-weighted mean concentrations decreased by 17–43%. The largest decreases occur in China in February with concentration decreases exceeding 10 parts per billion by volume (ppbv) and substantial decreases persisting in eastern urban areas through April. Thus these lockdown measures temporarily bolstered the decreasing trends across North America⁴² and Europe²⁵ over the last two decades and in China since 2012⁴³, owing to technological advances in vehicles and power generation, while temporarily buffering changes from increasing energy demands in India and the Middle East^{40,44,45}. NO₂ increases in April 2020 in central China (Chengdu and Chongqing) because lockdowns began lifting during this time.

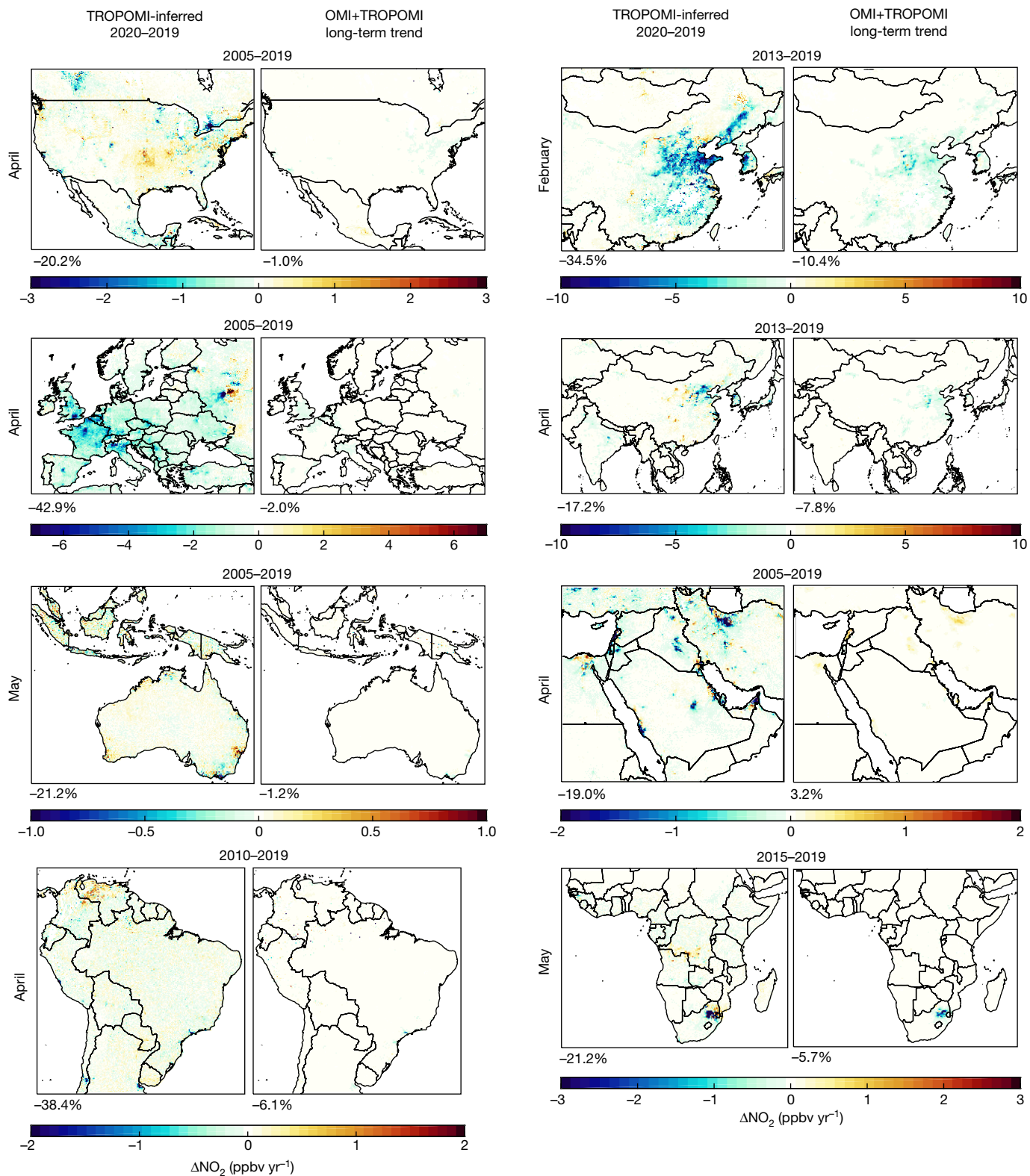


Fig. 3 | Changes in ground-level NO₂ during lockdowns. Left in each pair of images, TROPOMI-derived monthly mean NO₂ differences from 2020–2019 at approximately 1 × 1 km². Right, OMI+TROPOMI-derived NO₂ trends. Annual mean long-term trends are corrected for seasonal variation. The time periods

for trend calculations in each region were chosen to reflect the most recent years where a consistent trend is observed and are indicated above the maps. Value under each panel represents population-weighted mean difference for the given region.

Figure 3 shows maps of long-term NO₂ trends for context. In most regions, the observed changes during COVID-19 restrictions exceed the expected year-to-year differences observed in the long-term trends (Table 1). 2020–2019 population-weighted mean

concentration changes are lower than long-term trends by factors of 17 ± 7 in North America, 19 ± 2 Europe, of 2.9 ± 0.6 in Africa and the Middle East, of 3.6 ± 0.6 in Asia, 8 ± 7 in South America, and 2 ± 2 in Oceania.

Table 1 | TROPOMI-derived, population-weighted ground-level NO₂ data

Country/region	Month with greatest 2020–2019 change	Monthly population-weighted mean NO ₂ concentration 2019 (ppbv)	Monthly population-weighted mean 2020–2019 difference (ppbv)	Expected 2020–2019 change from meteorology (ppbv)	Long-term trend in population-weighted NO ₂ ^a (ppbv/year)	Ratio of 2020–2019 difference to long-term trend (years)
China ^b	January	9.5±0.3	-2.7±0.3	0.057±0.03	-0.8±0.1	3.4±0.6
India ^b	June	0.96±0.06	-0.29±0.03	-0.062±0.002	0.017±0.005	na
USA	March	3.0±0.1	-0.40±0.08	-0.12±0.01	-0.119±0.009	3.4±0.7
Indonesia ^b	June	1.24±0.04	-0.3±0.3	-0.031±0.007	-0.016±0.006	20±20
Brazil ^c	April	1.01±0.04	-0.3±0.3	-0.15±0.01	-0.064±0.007	5±4
Bangladesh ^b	April	0.82±0.05	-0.24±0.09	-0.18±0.01	0.026±0.006	na
Mexico	May	2.75±0.06	-0.68±0.07	0.01±0.01	0.095±0.006	na
Russia	April	4.18±0.07	-1.4±0.2	-0.39±0.02	-0.074±0.003	19±3
Japan ^b	April	4.0±0.3	-1.9±0.2	-0.19±0.02	-0.24±0.04	8±2
Egypt ^d	May	3.1±0.1	-0.4±0.2	-0.03±0.01	-0.25±0.09	1.4±0.9
Iran ^d	April	2.76±0.07	-0.5±0.7	0.080±0.008	-0.12±0.02	4±6
Turkey ^d	April	4.23±0.08	-1.5±0.7	0.17±0.03	0.135±0.007	na
Germany	March	7.95±0.3	-2.7±0.4	-0.77±0.01	-0.12±0.01	23±4
Thailand ^b	March	1.34±0.08	-0.25±0.03	-0.052±0.008	-0.003±0.008	100±200
France	April	4.76±0.03	-3.1±0.1	-0.117±0.008	-0.168±0.009	19±1
United Kingdom	April	6.42±0.03	-2.8±0.1	-0.19±0.02	-0.43±0.01	6.7±0.3
Italy	February	10.9±0.3	-2.8±0.3	-2.84±0.05	-0.37±0.02	8±1
South Africa ^d	May	7.7±0.1	-2.7±0.3	-0.06±0.02	-0.4±0.2	7±3
Spain	April	3.16±0.04	-2.1±0.1	-0.113±0.006	-0.169±0.009	12.6±0.9
Argentina ^c	April	1.63±0.07	-0.8±0.7	-0.32±0.02	-0.08±0.01	11±10
Africa ^d	May	0.66±0.02	-0.15±0.02	-0.012±0.001	-0.051±0.007	2.9±0.6
Asia ^b	March	3.0±0.1	-0.70±0.05	0.002±0.001	-0.19±0.03	3.6±0.6
East Asia ^b	February	6.4±0.1	-1.86±0.02	-0.068±0.001	-0.55±0.06	3.4±0.4
South Asia ^b	June	0.98±0.06	-0.28±0.03	-0.044±0.001	0.015±0.006	na
Europe	April	3.87±0.02	-1.67±0.08	-0.096±0.001	-0.090±0.007	19±2
West Europe	April	4.52±0.02	-2.08±0.07	-0.115±0.001	-0.163±0.009	12.8±0.9
Central Europe	April	2.86±0.05	-1.0±0.2	0.013±0.001	0.053±0.005	na
East Europe	April	3.43±0.03	-1.40±0.06	-0.167±0.001	-0.049±0.004	29±2
North America	April	2.41±0.07	-0.5±0.1	-0.105±0.001	-0.029±0.008	17±7
Oceania	May	1.59±0.09	-0.2±0.1	-0.024±0.001	-0.086±0.005	2±2
South America ^c	April	1.11±0.05	-0.4±0.4	-0.022±0.001	-0.056±0.007	8±7
Global (country level)	April	1.5±0.2	-0.53±0.06	-0.050±0.010	-0.04±0.01	15±4
Global (population-weighted)	April	2.2±0.5	-0.52±0.08	-0.06±0.04	-0.10±0.05	5±3

Countries with largest populations and annual mean population-weighted NO₂ concentrations greater than 1 ppbv are shown for months with the greatest 2020–2019 difference and strict lockdown conditions (stringency index >20), sorted by population. Regional and global data also shown.

^aSatellite-inferred annual mean ground-level NO₂ trends are scaled by the ratio of the 2019 monthly mean to the annual mean to account for seasonality.

Long-term country-level trends are calculated for 2005–2019, except for countries/regions in:

^bAsia: 2013–2019.

^cSouth America: 2011–2019.

^dAfrica and the Middle East: 2015–2019.

na, Ratio of 2020–2019 difference to long-term trend not calculated when one value is positive and one is negative.

Meteorological differences are calculated with the GEOS-Chem chemical transport model using emission inventories that do not include changes that occurred owing to COVID-19 lockdown policies but do reflect meteorological changes. Supplementary Fig. 10 shows TROPOMI-derived changes at 2.0° × 2.5° resolution for comparisons with simulated values at the same resolution. Population-weighted NO₂ concentration changes due to meteorology in Asia, Europe, South America, Africa and the Middle East are a factor of 2–6 smaller than observed; thus, meteorology alone cannot explain the observed decreases. Concentration increases in the central USA, as noted in other studies¹⁰, do not appear to be meteorologically driven and may be due to changes in biogenic NO_x sources.

Supplementary Fig. 11 shows the ratio of population-weighted January–June monthly mean NO₂ concentrations in 2020 to 2019 across selected regions. Most regions have the largest decrease in NO₂ in April when lockdown conditions were strongest (the global mean COVID restriction stringency index (defined in Methods) reached a maximum of 0.79 on 18 April), apart from China, where lockdowns were initiated in January. In most regions, 2020 NO₂ concentrations return towards pre-lockdown values in May or June owing to relaxing travel restrictions (30 June global mean stringency index, 0.60) as well as increasing soil, lightning and biomass-burning emissions that lessen the sensitivity of ambient NO₂ to anthropogenic emissions.

City- and country-level NO₂ changes

The fine resolution of our satellite-derived ground-level NO₂ dataset enables the assessment of larger changes in NO₂ concentrations from 2020–2019 evident at the city level. We calculate changes in TROPOMI-observed monthly mean ground-level NO₂ from 2020–2019 over 215 major cities (the ten most populous cities in each country with a population greater than 1 million) for the month with the greatest monthly mean lockdown stringency index, compared with expected changes due to meteorology and long-term trends (Supplementary Table 1). Most cities have TROPOMI-derived NO₂ decreases that cannot be explained by changes due to meteorology alone. For example, satellite-derived NO₂ concentrations in Beijing decreased by 45% in March, despite meteorological conditions favourable to increased NO₂. Jakarta, Manila, Istanbul, Los Angeles and Buenos Aires among others had decreased NO₂, despite similarly unfavourable meteorological conditions. Some cities, including Moscow, Tokyo, London, New York, Toronto and Delhi, had meteorological conditions that would have led to NO₂ decreases regardless of emission changes, but observed concentration changes exceeded the expected meteorological change.

Consistent analysis of individual cities as enabled by this dataset reveals a mean observed decrease of $32 \pm 2\%$ for these 215 cities. The mean expected meteorologically driven change was $-1 \pm 1\%$ and the mean expected change owing to long-term trends was a decrease of $1.4 \pm 0.4\%$. Supplementary Fig. 12 shows these reductions to be consistent with those found in 381 ground-monitor values from 79 studies³⁴ ($32 \pm 2\%$). Of the 215 cities included here, 65 are in countries that did not have ground-monitoring data available for previous studies. Notably, the 65 cities without monitors are largely in lower-income countries of Africa and southeast Asia. The average gross national income per capita for unmonitored countries is US\$7,100, compared to US\$25,000 for monitored countries, illustrating the potential of satellite-derived ground-level concentrations for providing information about lower-income regions. In summary, the observed decreases in NO₂ across more than 200 cities worldwide were generally driven by COVID-19 lockdowns, with locally varying modulation by meteorology and business-as-usual changes.

Table 1 shows monthly mean country-level population-weighted NO₂ concentrations, changes during COVID-19 lockdown restrictions, meteorological effects and long-term trends for the month with the greatest 2020–2019 change. Meteorological effects were generally minor at the national and regional scale. Multi-year trends provide context for the scale of the changes observed during COVID-19 lockdowns. The decrease in March NO₂ concentrations in the USA from 2019 to 2020 was equivalent to four years of long-term NO₂ reductions. Similarly, changes in NO₂ during COVID-19 lockdowns were equivalent to greater than three years of reductions in China, and up to 23 years in Germany. Globally, the April 2020 population-weighted NO₂ concentration was 0.53 ± 0.06 ppbv lower than in April 2019, equivalent to 15 ± 4 years of global NO₂ reductions.

NO₂ as a lockdown indicator

The relationship between this satellite-derived ground-level NO₂ dataset and lockdown stringency provides supporting evidence for the impact of travel restrictions (Supplementary Fig. 13). The ratio of population-weighted mean observed NO₂ in 2020 to 2019 was calculated for each country and each month from January to June. The 2020/2019 NO₂ ratio in countries with the strictest lockdown (monthly minimum stringency indices greater than the 75th percentile) was $29 \pm 3\%$ lower than for countries with the weakest lockdowns (monthly median stringency indices less than the 25th percentile). Maximum and median ratios were also lower for countries with strict lockdowns. Both distributions have similar variability (standard deviations 0.02 and 0.03) which demonstrates similar interannual variability due to

meteorology for both sets. When focusing on only the month with the strictest lockdown for each country, changes in population-weighted NO₂ are correlated with lockdown intensity, with changes in countries with strict lockdowns (average decrease 43% if lockdown index >80) more than three times as large as in those with weaker lockdowns (12% if lockdown index <40).

This relationship suggests that changes in satellite-derived NO₂ concentrations offer observational information on the spatial distribution of lockdown effects that is not available through policy-based stringency indices. For example, although the policy-based stringency index in most cases provides a single value for a country, city-level NO₂ concentration decreases in India are in the range 30–84%, reflecting variability in local mobility restrictions, emissions sources, and their sensitivity to lockdowns. Supplementary Fig. 14 explores the sensitivity of NO₂ concentrations to emissions from the transportation and electricity sectors in India, China and the USA by examining the distribution of changes in NO₂ concentration at the 20 largest population centres and 20 largest fossil fuel-burning power plants in each country. All countries have substantial NO₂ decreases in cities, but the sensitivities vary in areas associated with the electricity sector, with decreasing concentrations near power plants in India (mean change $-35 \pm 4\%$) and China ($-28 \pm 8\%$) but insignificant changes in the USA ($-4 \pm 8\%$). Observed NO₂ changes at these power plants exceed expected changes from meteorology alone ($-8 \pm 2\%$, $-1 \pm 4\%$ and $-1 \pm 3\%$ in India, China and the USA, respectively). Although variability between power plants reflects a mix of regionally varying factors, including meteorology, electricity demand, fuel type and plant-specific emission controls, as well as changes in nearby emissions from other sectors including transportation, these differences indicate a sensitivity of local air quality to activity restrictions affecting the energy sector.

Examining geographic differences in satellite-derived NO₂ concentrations within metropolitan regions is also informative. For example, variability between emission sources is apparent around the city of Atlanta, Georgia, USA (Supplementary Fig. 15). The population-weighted NO₂ concentration in Atlanta and the surrounding region dropped by 28% from April 2019 to 2020, but with substantial spatial variability in the observed change. The greatest NO₂ decreases are found near a large coal-powered electricity plant to the southeast of the city, with significant changes near another plant to the northwest. Decreases were also larger near the Hartsfield–Jackson International Airport—reflecting the dramatic slowdown in air travel—and over suburban regions to the west and northeast of the city centre, than in the downtown core. Supplementary Fig. 15 also demonstrates the range of NO₂ changes experienced by the local population. Over 1.2 million people live in regions where NO₂ decreases exceeded 40%, whereas nearly 1 million people experienced decreases of 10% or less. Similar heterogeneity in population exposure exists in other major cities, as demonstrated by Supplementary Fig. 16. For example, a subset of over 1 million people in the Paris metropolitan area experienced NO₂ decreases of 75% (4.5 ppbv) or more (10th-percentile exposure), whereas another similar-sized subset experienced changes of 23% (0.6 ppbv) or less (90th-percentile exposure). Of the cities examined here, 68 had an interquartile range in population exposure change during lockdowns of 20 percentage points or larger, 22 of which were unmonitored cities. Studies have found that NO₂ changes during lockdowns varied among socioeconomic, ethnic and racial groups in US cities⁴⁶, and thus the variability in other major cities observed here suggest similar disparities may occur elsewhere. The heterogeneity of NO₂ changes demonstrates the need for the finely resolved information on lockdown effects that is offered by satellite observations.

We find that using this satellite-derived NO₂ dataset as an observational proxy for lockdown conditions is also useful for identifying links between lockdown-driven emission changes and secondary pollutants. For example, several studies have found little to no change in fine particulate matter (PM_{2.5}) during lockdowns as meteorology, long-range

transport and nonlinear chemistry complicate the relationship between $PM_{2.5}$ and NO_x emissions^{47,48}. A challenge in these studies has been limited observational information on the local lockdown intensity. Recent work examining 2020–2019 changes in satellite-derived $PM_{2.5}$ concentrations found that lockdown-driven decreases in $PM_{2.5}$ concentration can be identified by separating the meteorological effects from emissions effects using chemical transport modelling and focusing on regions with the greatest sensitivity to emission reductions⁴⁹. Here we examine that same satellite-derived $PM_{2.5}$ dataset using TROPOMI-derived ground-level NO_2 concentrations to identify the regions where $PM_{2.5}$ concentrations are most likely associated with lockdowns or sensitive to NO_x emissions. Supplementary Fig. 17 shows the distribution of changes in monthly mean $PM_{2.5}$ concentrations from 2020–2019 for China in February and North America and Europe in April. Regions with the largest 2020–2019 NO_2 concentration decreases (90th percentile) are considered to be those with significant NO_x emission reductions. Population-weighted mean $PM_{2.5}$ concentrations decreased overall; however, regions with the largest NO_2 decreases experienced greater local changes in $PM_{2.5}$ concentration in China and to a lesser extent in North America, indicating that the sensitivity of $PM_{2.5}$ to changing NO_x emissions can be inferred. The year-to-year variability of $PM_{2.5}$ concentrations in Europe is similar regardless of changes in NO_2 , indicating a greater role of meteorology or transport on $PM_{2.5}$ in this region and period. These results are consistent with previous findings when using chemical transport modelling to identify regions where local emissions are important⁴⁹. Thus, the observational proxy on lockdown conditions offered by these satellite-derived surface NO_2 concentrations offers spatially resolved information to identify where $PM_{2.5}$ and NO_2 (and by proxy, NO_x emissions) are most strongly coupled.

Implications

The pronounced decreases in ground-level NO_2 found here for over 200 cities worldwide during COVID-19 lockdowns are a culmination of recent advancements in techniques for estimating ground-level NO_2 from satellite observations²⁷ alongside higher-resolution satellite observations from TROPOMI that allow for estimating high spatial resolution, short-term changes in NO_2 exposure. This method bridges the gap between monitor data (that measure ground-level air quality but have poor spatial representativeness) and satellite column data (that provide spatial distributions but are less representative of ground-level air quality). The ability to infer global ground-level NO_2 concentrations with sufficient resolution to assess individual cities and even within-city gradients is an important development in satellite remote-sensing instrumentation and algorithms. Additionally, these satellite-derived ground-level NO_2 concentrations offer information about unmonitored communities and populations that are underrepresented in studies focused on ground-monitor data. These cities are found to have different characteristics of NO_2 concentrations and changes during lockdowns that motivate the need for satellite observations in the absence of local ground monitoring. The changes in ground-level NO_2 due to COVID-19 lockdown restrictions, which exceed recent long-term trends and expected meteorologically driven changes, demonstrate the impact that policies that limit emissions can have on NO_2 exposure. This information has relevance to health impact assessment; for example, studies focused on ground-monitor data have indicated improvements in health outcomes related to improved air quality during lockdowns, including an estimated 780,000 fewer deaths and 1.6 million fewer paediatric asthma cases worldwide due to decreased NO_2 exposure²⁰. Our study demonstrates considerable spatial variability in lockdown-driven ground level NO_2 changes that does not necessarily correlate with population density, demonstrating probable uncertainties arising from extrapolating changes observed by ground monitors to estimate broad changes in population NO_2 exposure. Satellite-based ground-level NO_2 estimates provide high-resolution information on the spatial

distribution of NO_2 changes in 2020 that cannot be achieved through ground monitoring, particularly in regions without adequate ground monitoring, and should improve exposure estimates in future health studies. Additionally, ground-level concentrations from downscaled OMI observations provide the opportunity to contrast effects of past mitigation efforts on long-term NO_2 trends against the short-term changes resulting from more dramatic regulations, and a chance to improve studies of health outcomes related to long-term NO_2 exposure.

The strength of the links between observed changes in NO_2 concentration and lockdown stringency indicates that satellite-based ground-level NO_2 concentrations offer useful observational, spatially resolved information about lockdown conditions. This provides an observational metric for examining the efficacy of lockdown restrictions on restricting mobility for studies examining the spread of COVID-19. Here we exploited this information to illustrate the differing sensitivity of NO_2 concentrations to changes in various emission sources to lockdown restrictions. Future applications of these data could include examining socioeconomic drivers that impact this variability within and between countries. Comparisons between satellite-derived ground-level NO_2 and $PM_{2.5}$ also indicate the utility of these data as an observational proxy for identifying regions where secondary pollutants such as $PM_{2.5}$ or ozone are more likely to be sensitive to NO_x emissions; these links are otherwise difficult to trace without the use of chemical transport models⁵⁰.

These data offer information to improve NO_2 -exposure estimates, to examine exposure trends, and subsequently estimate changes in health burden. These developments provide an excellent opportunity for advances in air quality health assessment in relation to NO_2 and its combustion-related air pollutant mixture.

Online content

Any methods, additional references, Nature Research reporting summaries, source data, extended data, supplementary information, acknowledgements, peer review information; details of author contributions and competing interests; and statements of data and code availability are available at <https://doi.org/10.1038/s41586-021-04229-0>.

1. GBD 2019 Risk Factors Collaborators. Global burden of 87 risk factors in 204 countries and territories, 1990–2019: a systematic analysis for the Global Burden of Disease Study 2019. *Lancet* **396**, 1223–1249 (2020).
2. Pannullo, F. et al. Quantifying the impact of current and future concentrations of air pollutants on respiratory disease risk in England. *Environ. Health* **16**, 29 (2017).
3. Tao, Y., Mi, S., Zhou, S., Wang, S. & Xie, X. Air pollution and hospital admissions for respiratory diseases in Lanzhou, China. *Environ. Pollut.* **185**, 196–201 (2014).
4. Zeng, W. et al. Association between NO_2 cumulative exposure and influenza prevalence in mountainous regions: a case study from southwest China. *Environ. Res.* **189**, 109926 (2020).
5. Anenberg, S. C. et al. Estimates of the global burden of ambient $PM_{2.5}$, ozone, and NO_2 on asthma incidence and emergency room visits. *Environ. Health Perspect.* **126**, 107004 (2018).
6. Achakulwisut, P., Brauer, M., Hystad, P. & Anenberg, S. C. Global, national, and urban burdens of paediatric asthma incidence attributable to ambient NO_2 pollution: estimates from global datasets. *Lancet Planet. Health* **3**, e166–e178 (2019).
7. Hamra, G. B. et al. Lung cancer and exposure to nitrogen dioxide and traffic: a systematic review and meta-analysis. *Environ. Health Perspect.* **123**, 1107–1112 (2015).
8. Brook, J. R. et al. Further interpretation of the acute effect of nitrogen dioxide observed in Canadian time-series studies. *J. Expo. Sci. Environ. Epidemiol.* **17**, S36–S44 (2007).
9. Crouse, D. L. et al. Within-and between-city contrasts in nitrogen dioxide and mortality in 10 Canadian cities; a subset of the Canadian Census Health and Environment Cohort (CanCHEC). *J. Expo. Sci. Environ. Epidemiol.* **25**, 482–489 (2015).
10. Goldberg, D. L. et al. Disentangling the impact of the COVID-19 lockdowns on urban NO_2 from natural variability. *Geophys. Res. Lett.* **47**, e2020GL089269 (2020).
11. Biswal, A. et al. COVID-19 lockdown induced changes in NO_2 levels across India observed by multi-satellite and surface observations. *Atmos. Chem. Phys.* **21**, 5235–5251 (2021).
12. Koukouli, M.-E. et al. Sudden changes in nitrogen dioxide emissions over Greece due to lockdown after the outbreak of COVID-19. *Atmos. Chem. Phys.* **21**, 1759–1774 (2021).
13. Field, R. D., Hickman, J. E., Geogdzhayev, I. V., Tsigaridis, K. & Bauer, S. E. Changes in satellite retrievals of atmospheric composition over eastern China during the 2020 COVID-19 lockdowns. Preprint at <https://doi.org/10.5194/acp-2020-567> (2020).
14. Bauwens, M. et al. Impact of coronavirus outbreak on NO_2 pollution assessed using TROPOMI and OMI observations. *Geophys. Res. Lett.* **47**, e2020GL087978 (2020).

15. Liu, F. et al. Abrupt decline in tropospheric nitrogen dioxide over China after the outbreak of COVID-19. *Sci. Adv.* **6**, eabc2992 (2020).
16. Prunet, P., Lezeaux, O., Camy-Peyret, C. & Thevenon, H. Analysis of the NO₂ tropospheric product from S5P TROPOMI for monitoring pollution at city scale. *City Environ. Interact.* **8**, 100051 (2020).
17. Shi, X. & Brasseur, G. P. The response in air quality to the reduction of Chinese economic activities during the COVID-19 outbreak. *Geophys. Res. Lett.* **47**, e2020GL088070 (2020).
18. Ropkins, K. & Tate, J. E. Early observations on the impact of the COVID-19 lockdown on air quality trends across the UK. *Sci. Total Environ.* **754**, 142374 (2021).
19. Fu, F., Purvis-Roberts, K. L. & Williams, B. Impact of the COVID-19 pandemic lockdown on air pollution in 20 major cities around the world. *Atmosphere* **11**, 1189 (2020).
20. Venter, Z. S., Aunan, K., Chowdhury, S. & Lelieveld, J. COVID-19 lockdowns cause global air pollution declines. *Proc. Natl Acad. Sci.* **117**, 18984–18990 (2020).
21. Levy, I., Mihele, C., Lu, G., Narayan, J. & Brook, J. R. Evaluating multipollutant exposure and urban air quality: pollutant interrelationships, neighborhood variability, and nitrogen dioxide as a proxy pollutant. *Environ. Health Perspect.* **122**, 65–72 (2014).
22. Shi, Z. et al. Abrupt but smaller than expected changes in surface air quality attributable to COVID-19 lockdowns. *Sci. Adv.* **7**, eabd6696 (2021).
23. Liu, Q. et al. Spatiotemporal changes in global nitrogen dioxide emission due to COVID-19 mitigation policies. *Sci. Total Environ.* **776**, 146027 (2021).
24. Lamsal, L. N. et al. Ground-level nitrogen dioxide concentrations inferred from the satellite-borne Ozone Monitoring Instrument. *J. Geophys. Res.* **113**, D16308 (2008).
25. Geddes, J. A., Martin, R. V., Boys, B. L. & van Donkelaar, A. Long-term trends worldwide in ambient NO₂ concentrations inferred from satellite observations. *Environ. Health Perspect.* **124**, 281–289 (2016).
26. Gu, J. et al. Ground-level NO₂ concentrations over China inferred from the satellite OMI and CMAQ model simulations. *Remote Sens.* **9**, 519 (2017).
27. Cooper, M. J., Martin, R. V., McLinden, C. A. & Brook, J. R. Inferring ground-level nitrogen dioxide concentrations at fine spatial resolution applied to the TROPOMI satellite instrument. *Environ. Res. Lett.* **15**, 104013 (2020).
28. Levelt, P. F. et al. The Ozone Monitoring Instrument: overview of 14 years in space. *Atmos. Chem. Phys.* **18**, 5699–5745 (2018).
29. Levelt, P. F. et al. The Ozone Monitoring Instrument. *IEEE Trans. Geosci. Remote Sens.* **44**, 1093–1100 (2006).
30. Veeffkind, J. P. et al. TROPOMI on the ESA Sentinel-5 Precursor: a GMES mission for global observations of the atmospheric composition for climate, air quality and ozone layer applications. *Remote Sens. Environ.* **120**, 70–83 (2012).
31. Goldberg, D. L., Anenberg, S., Mohegh, A., Lu, Z. & Streets, D. G. TROPOMI NO₂ in the United States: a detailed look at the annual averages, weekly cycles, effects of temperature, and correlation with PM_{2.5}. Preprint at <https://doi.org/10.1002/essoar.10503422.1> (2020).
32. Dix, B. et al. Nitrogen oxide emissions from US oil and gas production: recent trends and source attribution. *Geophys. Res. Lett.* **47**, e2019GL085866 (2020).
33. Schenkeveld, V. M. E. et al. In-flight performance of the Ozone Monitoring Instrument. *Atmos. Meas. Tech.* **10**, 1957–1986 (2017).
34. Gkatzelis, G. I. et al. The global impacts of COVID-19 lockdowns on urban air pollution: a critical review and recommendations. *Elem. Sci. Anthr.* **9**, 00176 (2021).
35. Benitez-García, S.-E., Kanda, I., Wakamatsu, S., Okazaki, Y. & Kawano, M. Analysis of criteria air pollutant trends in three Mexican metropolitan areas. *Atmosphere* **5**, 806–829 (2014).
36. Duncan, B. N. et al. A space-based, high-resolution view of notable changes in urban NO_x pollution around the world (2005–2014). *J. Geophys. Res.* **121**, 976–996 (2016).
37. Bari, M. & Kindziarski, W. B. Fifteen-year trends in criteria air pollutants in oil sands communities of Alberta, Canada. *Environ. Int.* **74**, 200–208 (2015).
38. Zheng, B. et al. Trends in China's anthropogenic emissions since 2010 as the consequence of clean air actions. *Atmos. Chem. Phys.* **18**, 14095–14111 (2018).
39. Georgoulias, A. K., van der, A. R. J., Stammes, P., Boersma, K. F. & Eskes, H. J. Trends and trend reversal detection in 2 decades of tropospheric NO₂ satellite observations. *Atmos. Chem. Phys.* **19**, 6269–6294 (2019).
40. Krotkov, N. A. et al. Aura OMI observations of regional SO₂ and NO₂ pollution changes from 2005 to 2015. *Atmos. Chem. Phys.* **16**, 4605–4629 (2016).
41. Hilboll, A., Richter, A. & Burrows, J. P. NO₂ pollution over India observed from space – the impact of rapid economic growth, and a recent decline. Preprint <https://doi.org/10.5194/acp-2017-101> (2017).
42. Zhang, R. et al. Comparing OMI-based and EPA AQS in situ NO₂ trends: towards understanding surface NO_x emission changes. *Atmos. Meas. Tech.* **11**, 3955–3967 (2018).
43. Lin, N., Wang, Y., Zhang, Y. & Yang, K. A large decline of tropospheric NO₂ in China observed from space by SNPP OMI. *Sci. Total Environ.* **675**, 337–342 (2019).
44. Barkley, M. P. et al. OMI air-quality monitoring over the Middle East. *Atmos. Chem. Phys.* **17**, 4687–4709 (2017).
45. Vohra, K. et al. Long-term trends in air quality in major cities in the UK and India: a view from space. *Atmos. Chem. Phys.* **21**, 6275–6296 (2021).
46. Kerr, G. H., Goldberg, D. L. & Anenberg, S. C. COVID-19 pandemic reveals persistent disparities in nitrogen dioxide pollution. *Proc. Natl Acad. Sci.* **118**, e2022409118 (2021).
47. Le, T. et al. Unexpected air pollution with marked emission reductions during the COVID-19 outbreak in China. *Science* **369**, 702–706 (2020).
48. Chen, L.-W. A., Chien, L.-C., Li, Y. & Lin, G. Nonuniform impacts of COVID-19 lockdown on air quality over the United States. *Sci. Total Environ.* **745**, 141105 (2020).
49. Hammer, M. S. et al. Effects of COVID-19 lockdowns on fine particulate matter concentrations. *Sci. Adv.* **7**, eabg7670 (2021).
50. Keller, C. A. et al. Global impact of COVID-19 restrictions on the surface concentrations of nitrogen dioxide and ozone. *Atmos. Phys. Chem.* **21**, 3555–3592 (2021).

Publisher's note Springer Nature remains neutral with regard to jurisdictional claims in published maps and institutional affiliations.



Open Access This article is licensed under a Creative Commons Attribution 4.0 International License, which permits use, sharing, adaptation, distribution and reproduction in any medium or format, as long as you give appropriate credit to the original author(s) and the source, provide a link to the Creative Commons license, and indicate if changes were made. The images or other third party material in this article are included in the article's Creative Commons license, unless indicated otherwise in a credit line to the material. If material is not included in the article's Creative Commons license and your intended use is not permitted by statutory regulation or exceeds the permitted use, you will need to obtain permission directly from the copyright holder. To view a copy of this license, visit <http://creativecommons.org/licenses/by/4.0/>.

© The Author(s) 2022

Methods

Data

We use tropospheric NO₂ columns from the OMI (NASA Standard Product version 4)⁵¹ and TROPOMI^{52,53} satellite instruments. Both instruments measure solar backscatter radiation in the ultraviolet–visible (UV–vis) spectral bands on sun-synchronous orbits with local overpass times around 1:30 p.m. TROPOMI observations from April 2018–October 2020 are used to examine near-term NO₂, and OMI observations from January 2005–December 2019 are used to examine long-term trends. Observations with retrieved cloud fractions greater than 0.1 or flagged as poor quality or snow-covered (that is, TROPOMI quality assurance flag <0.75) are excluded. Although the resolution of TROPOMI observations is 3.5 × 5.5 km², several studies have demonstrated that oversampling techniques can provide accurate NO₂ maps at 1 × 1 km² resolution when averaging over a one-month period^{31,32,54}. An area-weighted oversampling technique^{55,56} is used to map daily satellite NO₂ column observations from TROPOMI onto a -0.01° × 0.01° (-1 × 1 km²) resolution grid and from OMI to a 0.1° × 0.125° (-10 × 10 km²) grid, as these resolutions balance the need of fine resolution for observing fine-scale structure and of minimizing the effects of sampling biases and noise in the observations. Supplementary Fig. 8 provides further evidence that a one-month period provides sufficient observations for a 1 × 1 km² map as the agreement between TROPOMI-derived surface concentrations and in situ observations does not deteriorate when the sampling period is reduced from one year to one month. Additionally, we compared 2019 monthly mean concentration estimates with the 2019 annual mean and find high correlation ($r = 0.90$), indicating similar spatial variability. We correct for sampling biases in the satellite records due to persistent cloudy periods or surface snow cover using a correction factor calculated with the GEOS-Chem chemical transport model described below by sampling the GEOS-Chem-simulated monthly or annual mean column densities to match the satellite.

We use hourly ground-level NO₂ measurements from monitors to constrain and evaluate the satellite-based estimates. Observations from the US Environmental Protection Agency Air Quality System (https://aqz.epa.gov/aqzweb/documents/data_mart_welcome.html) over the continental USA from 2005–2020, Environment and Climate Change Canada's National Air Pollution Surveillance Program (<http://maps-cartes.ec.gc.ca/rnspa-naps/data.aspx>) from 2005–2019, European Environment Agency (<https://aqportal.discomap.eea.europa.eu/index.php/users-corner/>) from 2005–2020, National Air Quality Monitoring Network in China from 2015–2020 were (obtained from <https://quotsoft.net/air>) were used. European monitors classified as near-road are excluded. Monthly and annual mean concentrations at each site are calculated by averaging hourly observations between 13:00–15:00 h (corresponding to satellite overpass times) and corrected for the known overestimate in regulatory measurements due to interference of other reactive nitrogen species following Lamsal et al.²⁴.

To examine the relationship between COVID-19 lockdown policies and ground-level NO₂ concentrations, we use the Oxford COVID-19 Government Response Tracker (OxCGRT, <https://www.bsg.ox.ac.uk/research/research-projects/coronavirus-government-response-tracker/#data>). OxCGRT provides a daily country-level policy 'stringency index' ranging from 0–100 that is based on containment and closure policies (for example, school and workplace closures, stay-at-home orders, gathering restrictions). We also use population density data from the Center for International Earth Science Information Network for the available years of 2005, 2010, 2015 and 2020, and linearly interpolate for other years (<https://doi.org/10.7927/H4JW8BX5>).

Inferring ground-level concentrations from satellite column observations

Ground-level NO₂ concentrations are derived from TROPOMI NO₂ columns following the method developed in Cooper et al.²⁷.

This algorithm builds upon the method first developed by Lamsal et al.²⁴ which uses the GEOS-Chem-simulated relationship between ground-level and tropospheric column NO₂ concentrations. The updated algorithm uses the satellite-observed column densities and ground-monitor data as observational constraints on the shape of the boundary layer profile, reducing the sensitivity to model resolution and improving agreement between satellite-derived ground-level concentrations and in situ observations. Technical details on the application of this method as used here are available in the Supplementary Information.

For long-term trend analysis, we use more recent TROPOMI observations to provide fine-resolution spatial structure to the OMI-observed NO₂ columns following the method of Geddes et al.²⁵. Annual mean OMI NO₂ columns are gridded to 10 × 10 km² resolution and a median-value filter is applied to reduce noise. We smooth the two-year (April 2018–April 2020) mean TROPOMI NO₂ columns mapped at 1 × 1 km² resolution using a two-dimensional boxcar algorithm with an averaging window of 10 × 10 km² to match the resolution of the gridded OMI NO₂ columns. We then downscale the annual mean OMI NO₂ columns using the ratio of the 1 × 1 km² TROPOMI columns to the smoothed TROPOMI columns. The downscaled columns are then used to infer ground-level concentrations following the method used for TROPOMI. Supplementary Fig. 18 demonstrates the utility of this downscaling approach by comparing OMI-derived ground-level concentrations to those derived from the downscaled columns. When comparing 2020–2019 changes in monthly mean concentrations to long-term trends, trends in annual mean concentration are scaled by the ratio of the 2019 monthly mean to the 2019 annual mean to account for seasonality.

The GEOS-Chem chemical transport model version 11-01 is used here (<https://geos-chem.seas.harvard.edu/>) for NO₂ vertical profiles and to assess meteorological effects. GEOS-Chem simulates atmospheric chemistry and physics using a detailed HO_x–NO_x–VOC–O₃–aerosol chemical mechanism^{57,58} driven by meteorological data from the MERRA-2 Reanalysis of the NASA Global Modeling and Assimilation Office⁵⁹. A detailed description of the simulation is provided in Hammer et al.⁶⁰. We replace the a priori profile used in the retrieval with profiles simulated using the GEOS-Chem model to ensure consistency in vertical profile representation between TROPOMI, OMI, and GEOS-Chem. We simulate NO₂ profiles from January 2005–June 2020 at a horizontal resolution of 2° × 2.5°. Supplementary Fig. 19 shows results from tests using a simulation at 0.5° × 0.625° which was available over North America, Europe and Asia. Satellite-derived ground-level concentrations at -1 × 1 km² resolution were not sensitive to the resolution of the a priori information, consistent with Cooper et al.²⁷, and thus the 2° × 2.5° was used here for consistency across all regions.

Inferring country- and city-level NO₂ changes during COVID lockdowns

City-level monthly means are calculated from TROPOMI-derived concentrations at -1 × 1 km² resolution averaged over a 20 × 20 km² region surrounding the city. Meteorological effects are estimated using GEOS-Chem simulations at 2° × 2.5° resolution with consistent emissions in both years, downscaled to -1 × 1 km² resolution using the horizontal variability of TROPOMI-derived ground-level concentrations. Supplementary Fig. 20 demonstrates that GEOS-Chem simulations can represent meteorologically driven changes in NO₂ in pre-lockdown periods. Trends are defined over 2005–2019 for North America, Europe and Australia, 2015–2019 for Asia and Africa, and 2010–2019 for South America and scaled for seasonality.

Country-level population-weighted means, used to represent population NO₂ exposure, are calculated using concentrations at -1 × 1 km² resolution via:

$$\text{population-weighted mean} = \frac{\sum_{i=1}^{\text{grid boxes in country}} P_i x_i}{\sum_{i=1}^{\text{grid boxes in country}} P_i}, \quad (2)$$

where x_i is the NO₂ concentration and P_i is the population within a 1×1 -km² grid box.

Limitations and sources of uncertainty

Uncertainty values for country- and region-level population-weighted means (σ_{total}) represent the sum in quadrature of three main error sources:

$$\sigma_{\text{total}} = \sqrt{\sigma_{\text{pop-weighted}}^2 + \sigma_{\Omega_{\text{max}}}^2 + \sigma_{\text{AMF2020}}^2} \quad (3)$$

Uncertainty in population-weighted means ($\sigma_{\text{pop-weighted}}$) are estimated using a bootstrapping method⁶¹. Uncertainty in 2020 NO₂ estimates (σ_{AMF2020}) arises from the use of simulated profiles as a priori information for calculating satellite air mass factors and for informing the column-to-ground-level relationship, as these simulations use emission inventories that do not reflect changes resulting from COVID-19-related travel restrictions. Such errors may result in overestimating the fraction of columnar NO₂ near the surface, resulting in an overestimate in satellite-derived ground-level NO₂ concentrations and an underestimate of the 2020–2019 difference. We estimate σ_{AMF2020} by performing sensitivity studies where anthropogenic NO_x emissions were uniformly reduced by 50% to assess the effect of such emission errors on ground-level NO₂ estimates. Reducing anthropogenic NO_x emissions by 50% led to a 5% change in monthly mean population weighted NO₂ concentrations in North America, Europe and Asia for March 2020. Aerosols can also contribute to uncertainty in air mass factor calculations, as a reduction in anthropogenic scattering aerosols during lockdowns may reduce air mass factors leading an underestimation of the NO₂ change^{62,63}. However, this is likely to be a minor source of uncertainty in estimated NO₂ changes due to lockdown, because aerosol concentration changes were small in most regions⁴⁹ and a reduction in aerosol concentration of 10% translates to an uncertainty in NO₂ of less than 5%⁶⁴. Additional uncertainty ($\sigma_{\Omega_{\text{max}}}$) may arise from the choice of the Ω_{max} parameter (described in the Supplementary Information), particularly in regions where there are insufficient ground-monitor data for constraining Ω_{max} . We estimate $\sigma_{\Omega_{\text{max}}}$ by evaluating the sensitivity of mean population-weighted NO₂ concentrations to a 20% change in Ω_{max} . Median country-level $\sigma_{\Omega_{\text{max}}}$ values are ~7%. Uncertainty values in trends are calculated by a weighted linear regression where annual mean concentrations are weighted by σ_{total} .

Although tests here indicate that satellite-derived ground-level NO₂ concentrations are insensitive to the resolution of the simulated data used in the algorithm, discontinuities can occur at the edges of simulation grid boxes. To quantify this uncertainty, we calculate the difference across the grid box boundaries in each region. In most regions the discontinuity is small (<0.5 ppbv in 92% of total cases, and in 98% of cases where NO₂ concentrations >2 ppbv) although can be larger in some cases (>2 ppbv in 0.02% of cases where NO₂ concentrations >2 ppbv, maximum of 4.5 ppbv).

The along-track resolution of TROPOMI observations changed from 7 km to 5.5 km in August 2019. This change may influence interannual comparisons, particularly with respect to the sub-grid downscaling of process which relies on the spatial structure observed by the satellite. To test the influence of this change, we perform a case study where annual mean surface concentrations over Asia are calculated using two different sub-grid scaling factors (v in equation S1 in the Supplementary Information) determined from one year of observations before and after the resolution change, with other variables held constant. The mean relative difference between the two tests was 9% for grid boxes with annual mean concentrations greater than 1 ppbv, with a change in regional population-weighted NO₂ concentrations of 3%. Greater sensitivity to observation resolution was evident in regions with larger NO₂ enhancements, although relative differences greater than 25% occur in fewer than 5% of grid boxes. These tests indicate that although the change in observation resolution may change some spatial gradients, the overall impact on population exposure estimates is small.

Uncertainty values presented above represent uncertainty in the conversion of satellite-observed slant columns into surface concentrations and do not represent systematic errors in the retrieval of slant columns from satellite-observed radiances (~10%), or errors in the air mass factor calculations (23–37%), both of which have been extensively examined in prior studies^{52,65}. Errors related to air mass factor calculations can be reduced by using higher-resolution inputs in air mass factor calculations^{66,67} and are partially mitigated here during the conversion of column densities to surface concentrations through the sub-grid parameterization²⁷.

Although we apply a scaling factor to correct for sampling biases due to persistent cloud cover or surface snow cover, biases in monthly mean calculations may persist if the sampling rate is sufficiently low, particularly for city-level calculations. Most of the cities examined in Supplementary Table 1 had sufficient sampling to allow for a robust monthly mean calculation (median sampling rate of 14 days per month for the months indicated in the table), except for two cities for which fewer than 5 days of observations per month were available for the given month in either 2019 or 2020 (labelled * in Supplementary Table 1). However, results from these cities were consistent with nearby, more frequently sampled cities, lending confidence to these results despite the lower sampling frequency.

This dataset represents substantial improvement over past satellite-derived ground-level NO₂ estimates, as the updated algorithm is less sensitive to model resolution and leverages higher-resolution satellite observations than previous estimates. However, limitations remain. There can be considerable fine-scale variability at scales finer than the 1×1 km² resolution used here that cannot be captured by the satellite observations^{68,69}. Additionally, ground-monitor data are used as a constraint in converting observed column densities to ground-level concentrations, and thus absolute concentration values are probably less accurate in time periods or regions where ground-monitor data are unavailable. However, these data are still useful for examining relative interannual variability or trend analysis. In combining OMI and TROPOMI observations we assume that the spatial gradients observed by TROPOMI in 2018–2020 can be applied to OMI for the entire 2005–2019 time series. New or disappearing point emission sources with small plume footprints may affect this assumption; however, past evaluations of similar assumptions have not found it to be a substantial error source²⁵. Additional errors in the column to ground-level conversion may occur in areas with substantial free tropospheric NO₂ sources such as aircraft emissions or lightning.

Data availability

TROPOMI-derived 2019 annual mean ground-level NO₂ concentrations developed here are available at <https://doi.org/10.5281/zenodo.5484305>. TROPOMI-derived January–June 2019 and 2020 concentrations are available at <https://doi.org/10.5281/zenodo.5484307>. Satellite-derived ground-level NO₂ concentrations for 2005–2019 used for trend analysis are available at <https://doi.org/10.5281/zenodo.5424752>. Satellite column data used here are available from the NASA Goddard Earth Sciences Data and Information Services Center (TROPOMI, <https://doi.org/10.5270/S5P-s4lfg54>; OMI, 10.5067/Aura/OMI/DATA2017). The GEOS-Chem model version used here is available at <https://doi.org/10.5281/zenodo.2658178>. Hourly ground-level NO₂ measurements from ground monitors in the USA are available from the US Environmental Protection Agency Air Quality System (https://aq5.epa.gov/aq5web/documents/data_mart_welcome.html), in Canada from Environment and Climate Change Canada's National Air Pollution Surveillance Program (<http://maps-cartes.ec.gc.ca/rnspa-naps/data.aspx>), in Europe from the European Environment Agency (<https://aqportal.discomap.eea.europa.eu/index.php/users-corner/>), and in China from <https://quotsoft.net/air>. COVID-19 lockdown policy information is provided by the Oxford COVID-19

Article

Government Response Tracker (<https://www.bsg.ox.ac.uk/research/research-projects/coronavirus-government-response-tracker#data>). Population distribution data are available from the Center for International Earth Science Information Network, <https://doi.org/10.7927/H4JW8BX5>. NO₂ changes during COVID-19 lockdowns from previous studies used for comparison here were compiled by Gkatzelis et al.³⁴ and are available at <https://covid-aqs.fz-juelich.de>. Gross National Income data were provided by World Bank, available at https://data.worldbank.org/indicator/ny.gnp.pcap.cd?year_high_desc=true.

Code availability

Code used to calculate surface NO₂ concentrations from satellite columns is available upon request. Some features in the displayed maps were produced using The Climate Data Toolbox for MATLAB⁷⁰.

51. Lamsal, L. N. et al. Ozone Monitoring Instrument (OMI) Aura nitrogen dioxide standard product version 4.0 with improved surface and cloud treatments. *Atmos. Meas. Tech.* **14**, 455–479 (2021).
52. van Geffen, J. et al. S5P TROPOMI NO₂ slant column retrieval: method, stability, uncertainties and comparisons with OMI. *Atmos. Meas. Tech.* **13**, 1315–1335 (2020).
53. Folkert Boersma, K. et al. Improving algorithms and uncertainty estimates for satellite NO₂ retrievals: results from the quality assurance for the essential climate variables (QA4ECV) project. *Atmos. Meas. Tech.* **11**, 6651–6678 (2018).
54. Goldberg, D. L. et al. Enhanced capabilities of TROPOMI NO₂: estimating NO_x from North American cities and power plants. *Environ. Sci. Technol.* **53**, 12594–12601 (2019).
55. Spurr, R. *Area-weighting Tessellation For Nadir-Viewing Spectrometers*. Internal Technical Note (Harvard-Smithsonian Center for Astrophysics, 2003).
56. Zhu, L. et al. Formaldehyde (HCHO) as a hazardous air pollutant: mapping surface air concentrations from satellite and inferring cancer risks in the United States. *Environ. Sci. Technol.* **51**, 5650–5657 (2017).
57. Bey, I. et al. Global modeling of tropospheric chemistry with assimilated meteorology: model description and evaluation. *J. Geophys. Res. Atmos.* **106**, 23073–23095 (2001).
58. Park, R. J., Jacob, D. J., Field, B. D., Yantosca, R. M. & Chin, M. Natural and transboundary pollution influences on sulfate-nitrate-ammonium aerosols in the United States: implications for policy. *J. Geophys. Res. Atmos.* **109**, D15204 (2004).
59. Rienecker, M. M. et al. MERRA: NASA's Modern-Era Retrospective Analysis for Research and Applications. *J. Clim.* **24**, 3624–3648 (2011).
60. Hammer, M. S. et al. Global estimates and long-term trends of fine particulate matter concentrations (1998–2018). *Environ. Sci. Technol.* **54**, 7879–7890 (2020).
61. Gatz, D. F. & Smith, L. The standard error of a weighted mean concentration—I. Bootstrapping vs other methods. *Atmos. Environ.* **29**, 1185–1193 (1995).
62. Chimot, J., Vlemmix, T., Veefkind, J. P., de Haan, J. F. & Levelt, P. F. Impact of aerosols on the OMI tropospheric NO₂ retrievals over industrialized regions: how accurate is the aerosol correction of cloud-free scenes via a simple cloud model? *Atmos. Meas. Tech.* **9**, 359–382 (2016).
63. Lin, J.-T. et al. Retrieving tropospheric nitrogen dioxide from the Ozone Monitoring Instrument: effects of aerosols, surface reflectance anisotropy, and vertical profile of nitrogen dioxide. *Atmos. Chem. Phys.* **14**, 1441–1461 (2014).
64. Cooper, M. J., Martin, R. V., Hammer, M. S. & McLinden, C. A. An observation-based correction for aerosol effects on nitrogen dioxide column retrievals using the Absorbing Aerosol Index. *Geophys. Res. Lett.* **46**, 8442–8452 (2019).
65. Verhoelst, T. et al. Ground-based validation of the Copernicus Sentinel-5P TROPOMI NO₂ measurements with the NDACC ZSL-DOAS, MAX-DOAS and Pandonia global networks. *Atmos. Meas. Tech.* **14**, 481–510 (2021).
66. Laughner, J. L., Zare, A. & Cohen, R. C. Effects of daily meteorology on the interpretation of space-based remote sensing of NO₂. *Atmos. Chem. Phys.* **16**, 15247–15264 (2016).
67. Liu, S. et al. An improved air mass factor calculation for nitrogen dioxide measurements from the Global Ozone Monitoring Experiment-2 (GOME-2). *Atmos. Meas. Tech.* **13**, 755–787 (2020).
68. Judd, L. M. et al. Evaluating the impact of spatial resolution on tropospheric NO₂ column comparisons within urban areas using high-resolution airborne data. *Atmos. Meas. Tech.* **12**, 6091–6111 (2019).
69. Kharol, S. K. et al. Assessment of the magnitude and recent trends in satellite-derived ground-level nitrogen dioxide over North America. *Atmos. Environ.* **118**, 236–245 (2015).
70. Greene, C. A. et al. The Climate Data Toolbox for MATLAB. *Geochem. Geophys. Geosyst.* **20**, 3774–3781 (2015).

Acknowledgements This research was supported by Environment and Climate Change Canada and by the Canadian Urban Environmental Health Research Consortium. R.V.M. acknowledges support from NASA grants 80NSSC21K1343 and 80NSSC21K0508. We thank the OMI instrument team, and the OMI and TROPOMI teams for making NO₂ data publicly available.

Author contributions M.J.C. and R.V.M. designed the study. M.J.C. performed the analysis. M.S.H. performed GEOS-Chem model simulations and developed the PM_{2.5} data used here. P.F.L. and P.V. developed and provided the TROPOMI NO₂ data used here. L.N.L. and N.A.K. developed and provided the OMI NO₂ data used here. M.J.C. prepared the manuscript with contributions from R.V.M., M.S.H., P.F.L., P.V., L.N.L., N.A.K., J.R.B. and C.A.M.

Competing interests The authors declare no competing interests.

Additional information

Supplementary information The online version contains supplementary material available at <https://doi.org/10.1038/s41586-021-04229-0>.

Correspondence and requests for materials should be addressed to Matthew J. Cooper.

Peer review information Nature thanks the anonymous reviewers for their contribution to the peer review of this work.

Reprints and permissions information is available at <http://www.nature.com/reprints>.